

HEALTHCARE WASTE TREATMENT TECHNOLOGIES AND HEALTH IMPACTS OF WASTE MANAGEMENT

NESLI AYDIN

Faculty of Engineering, Department of Environmental Engineering, University of Karabuk,
Karabuk, Turkey

ABSTRACT

Healthcare waste includes all types of waste generated from human and veterinary medical activities (public or private), and encompasses diagnosis, prevention and curative treatments, research, and laboratory activities. The risks associated with healthcare waste and its management have gained attention across the world over last decades, and this has resulted in the increased recognition of the need for proper healthcare waste management. One of the most important factors in any type of waste management is undoubtedly safety and hence the basic safety for waste treatment workers on the site has to be guaranteed first. However, some waste treatment technologies, such as landfills and incinerators, release toxic substances which could cause significant health effects. There are two purposes of this paper: (1) to review the technologies available for the treatment of healthcare wastes (HCW) by indicating the major merits and drawbacks of them, and (2) to highlight adverse health implications of these technologies reported in the epidemiology to provide data and information regarding excess health risk of these technologies for further studies. If there are not a significant number of health effect investigations specifically related to HCW treatment and disposal, municipal waste will be used as a surrogate for healthcare waste.

Keywords: health effect, healthcare waste, incinerator, landfills, safety, toxic substances.

1 INTRODUCTION

Adoption of proper healthcare waste service became a priority for regulatory authorities since the destructive impacts of improper management of healthcare waste were figured out by the environmental agencies. However, it has recently been understood that treating wastes at incinerators or at landfills could have a high potential to cause health problems as these treatment methods release toxic substances [1, 2]. For example, studies on workers at incinerators and populations residing near to these plants have identified a wide range of associated health impacts. These studies give rise to great concerns regarding possible health impacts from incinerators, even though the number of studies is highly limited. For this reason, more sophisticated health surveillance of waste disposal and treatment facilities is still needed for a better understanding of these methods' health implications.

This paper presents some of the most common treatment methods utilized in the management of HCW. The methods discussed include: high- temperature technology (incineration), non-burn low-temperature technologies (autoclave, hydroclave and microwave) and landfill. Background information on the technologies is included to provide information to those who may not be familiar with the details of each alternative. This research aims to review:



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1. Recent technologies available for the treatment of HCW with their advantages and disadvantages by investigating a number of technology producers across Europe, the US and Canada or currently commercialised companies across the world.
2. Adverse health implications of these technologies reported in the epidemiology to provide data and information regarding excess health risk for further studies.

2 AVAILABLE TECHNOLOGIES FOR HEALTHCARE WASTES

Available technologies for HCW will be examined under two groups: (1) High-temperature (incineration) technology and (2) Non-burn/low-temperature alternative technologies. Since landfilling of HCW without pre-treatment is forbidden by the EU Landfill Directive [3], it is not involved as an option by itself, but reviewed in this section as it is one of the most common methods applied in developing countries.

Table 1 provides the HCW treatment and disposal methods across the world based on the analysis of available literature.

2.1 Landfilling

In developing countries, landfills are generally operated like an open dump [4]. In practice, HCW is dumped in the pits mixed with municipal wastes, and later burned [5]. Yong *et al.* [6] reported that in Nanjing Province, China, since a disposal cost mechanism had not been developed based on the market economics, higher disposal costs often encouraged some hospitals to dispose of their HCW by themselves.

2.2 High-temperature (incineration) technology

Incineration is a high-temperature dry oxidation process that converts the waste into residual ash and gases. It consists of a primary combustion chamber operating at 800–1000°C and a secondary chamber operating at 850–1100°C.

There are often two shortcomings regarding the use of incinerators in developing countries reported in the literature. Firstly, these incinerators are poorly designed and run inappropriately. Nemathaga [5] investigated an incinerator of Tshilidzini hospital in Limpopo Province and found out that the incinerator was generating high amounts of ash because of incomplete burning of the waste. Secondly, they require high investment, operation and maintenance costs along with costly emission control equipment [20].

2.3 Non-burn/Low-temperature alternative technologies

Currently available alternative technologies for the treatment of HCW require suitable land disposal facilities [19]. The main principle of these technologies is to render the waste “safe”. “Rendering safe” is defined by the Safe Management of Healthcare Waste Document published by the Department of Health in the UK DoH [21] to be applied to:

(1) Infectious waste: It demonstrates the ability to reduce the number of infectious organisms present in the waste to a level that no additional precautions are needed to protect workers or the public against infection by the waste; (2) Anatomical waste: It destroys anatomical waste such that it is no longer generally recognisable; (3) All clinical waste (including any equipment and sharps): They render all clinical waste unusable and unrecognisable; (4)

Table 1: Various countries and their common healthcare waste treatment methods.

Country	Method	Reference
Bangladesh (Dhaka City)	Dumping Autoclave	Hassan et al. [7]
Brazil (State of Rio Grande do Sul)	Incineration Autoclave	Da Silva et al. [8]
Denmark	Incineration Other Alternative Technologies	Bagge [9]
Germany	Incineration Autoclave	Hempen [10]
Greece (Central Macedonia)	Incineration Autoclave	Karagiannidis et al. [11]
India	Open burning Open dumping	Patil and Shekdar [12]
Iran (Fars Province)	Open dumping Incineration	Askarian et al. [13]
Libya	Open dumping Incineration	Sawalem et al. [14]
Nigeria (Ibadan)	Open dumping Incineration	Coker et al. [15]
Palestinian Territory	Open burning Thermal disinfection Incineration	Al-Khatib and Sato [16]
South Africa (Limpopo Province)	Incineration Autoclave Open dumping Landfill	Nemathaga et al. [5]
Sweden	Incineration Autoclave	Christiansson [17]
Turkey	Landfilling Incineration Autoclave	Ciplak [18]
UK	Incineration Alternative Technologies	Tudor et al. [19]

Medicinal waste: It destroys the component chemicals of chemical or medicinal and medically contaminated waste.

Following the same document, for infectious waste, the treatment must demonstrate, as a minimum, Level III criteria. For cultures of pathogenic microorganisms (pre-maceration or shredding is not appropriate for such wastes) it should show at least Level IV criteria provided by the US State and Territorial Association on Alternative Treatment Technologies (STAAT) guidelines [22].

2.4 Autoclaves

The autoclave consists of a metal cylindrical vessel which is surrounded by a steam jacket. Waste containers are loaded into a vessel on a cycle/batch base and are exposed to elevated temperature/pressure for a set time period (e.g. 121°C for 30 minutes, [23]). Steam is added into the system in order to maintain a prescribed temperature for a given period of time. The steam jacket reduces condensation in the vessel and thus reduces the loss of heat.

In practice, steam is supplied into the system via a boiler. Usually, boilers are heated by means of conventional fuels (such as gas, diesel, coal, or biomass) or they use electricity [24]. The selection of a proper boiler for the system is crucial in terms of having a sufficient amount and quality of steam to match the requirements of a system.

For verification, sufficient steam penetration and exposure time have occurred, biological (spores) or chemical indicators (colour-changing) are placed periodically in waste loads [24].

2.5 Hydroclaves

The hydroclave is basically a double-jacketed vessel with fragmenting paddles inside. After the door is closed, high-temperature steam is introduced to the outside jacket to heat the waste via the hot inner surface.

Although the basic principal on which autoclave and hydroclave is based is the same, there is a crucial difference between them in terms of steam recycling. In order for the standard autoclave to function, steam is injected into the sterilising vessel. This steam is then lost when the cycle ends. On the other side, in hydroclave, the steam is injected into the jacket, not into the vessel where the waste is sterilised and therefore the steam is never in contact with waste. This enables the hydroclave to reclaim some amount of steam back to the boiler [25]. However, one of the disadvantages of the hydroclave over the autoclave is that it takes more steam to heat up initially as it has to transfer the heat from the outer jacket into the vessel chamber through conduction. This initial high-energy requirement then diminishes for the continuing cycles.

The healthcare waste, which is treated by alternative treatment plants (autoclave, hydroclave or microwave) requires to be landfilled afterwards. Currently, there is no evidence in the literature reporting the formation of hazardous emissions due to treatment of waste by alternative treatments. However, the electricity, and the fuel oil required to run these plants indirectly cause emissions to be released.

2.6 Microwaves

Microwaves are electromagnetic waves with frequencies falling below the range for infrared waves and above the ultra-high frequency [4]. Its working principle is based on converting electrical energy into microwave energy. This microwave energy is used to produce steam from the moisture present in the HCW stream. Some systems apply low-frequency radio waves to inactivate microorganisms contained within the waste.

One of the disadvantages of microwave systems is their cost which might not be economically competitive compared to other technologies, especially in developing countries [26].

3 EPIDEMIOLOGIC STUDIES OF WASTE TREATMENT METHODS

Much of the current understanding of the health impacts of waste disposal is based on the application of epidemiological methodology as stated by Hester and Harrison [27]. In this

part of the study, the available epidemiological literature on the health effects among the workers in landfilling sites and incinerators was reviewed systematically to provide data and information regarding excess risk estimates for the worker's health.

The studies evaluated in this research were selected accordingly to the criteria proposed by Hester and Harrison [27]: (1) They have to be conducted in authorised incinerations or landfills; meaning that the ones considering open burning or unregulated disposal sites were disregarded; (2) They must provide some degree of consistency with other different epidemiological studies in terms of the types and significance of the outcomes; (3) They must have a theoretical basis in linking adverse health effects and exposure pathway and (4) They must have a basis for the effects, as indicated by actual measurements or examinations.

Giusti [28] categorised these studies into three groups:

(1) **Prospective Cohort Studies:** Two cohorts of people (exposed and non-exposed) who differ with respect to certain factors under study were followed over a period to determine how these factors affected rates of a certain outcome. This kind of study generally involves the collection and analysis of blood or tissue samples. For example: Unuvar *et al.* [29] conducted a survey to assess whether pregnant women were at risk of mercury intoxication due to fish consumption by taking blood samples from mothers and their new born babies. Mudge *et al.* [30] described the prevalence of inadequate energy and protein intake in older inpatients by screening consecutive patients admitted between November 2007 and March 2008 to the Royal Brisbane and Women's Hospital in Australia. Likewise Hoek *et al.* [31] examined the association between mortality and indicators of traffic-related air pollution in the Netherlands by investigating a random sample of 5000 people from 1986 to 1994.

On the other side, having too many repeated measurements and the selection of the measurement time points of cohort studies cause these studies to have an ad-hoc basis according to Tekle *et al.* [32] who pointed out the necessity of optimal design methods with a controlled budget for these studies.

(2) **Retrospective Case-Control Studies:** A case group of people who have already developed a specific disease, and a control group of healthy people are selected. Information on past exposure is collected retrospectively (generally via interviews with the participants).

These studies are relatively inexpensive compared to prospective cohort studies as (A) they involve smaller groups of people, (B) they do not generally require structured experiments, but are more prone to bias [28]. For instance: The study by Burke and Sawchuk [33] was based on 244 women who died from tuberculosis between 1874 and 1884. Some 12% of them had given birth within the year preceding their death. The study used the records in the local government death registries; and indicated that recent childbirth did not increase the risk of tuberculosis mortality among these women.

(3) **Cross-Sectional Studies:** They take account a specific group of the exposed population over a short period of time. They are 'cross sectional' because data is collected at one point in time. They can only be useful to generate hypotheses that can be tested later by more comprehensive studies; otherwise they might not be effective at distinguishing whether a particular disease developed before or after the group was exposed to a potential hazard as they do not look at time trends. There are a number of examples of cross-sectional studies in the literature as they are relatively cheap to carry out [34–37]

For the definition of the strength of the association between exposure to a potentially toxic substance and specific health effects in epidemiological studies, the ratio of the incidence of a disease in the exposed population to the incidence of the same disease in the non-exposed population is calculated; this is called "Relative Risk" (RR) or "Odd Risk" (OR). For instance,

if the RR is 6, the risk is six times higher (or an increase of 500%) in the exposed population than that in the non-exposed population.

The number of the studies satisfying the criteria set by Hester and Harrison [27] is very limited. Regarding mortality and morbidity among landfill workers, there is only one study: Gelberg [38] carried out a cross-sectional study to examine acute health effects among employees working for the New York City Department of Sanitation. Landfill workers reported a significantly higher prevalence of work-related respiratory (RR = 2.14), dermatologic (RR = 2.07), neurologic (RR = 1.89), gastrointestinal (RR = 1.26) and hearing problems (RR = 1.73), itching eyes (RR = 1.54) and sorethroat (RR = 2.26) than the controls.

Regarding the adverse health effects on incineration workers, Gustavsson [39] investigated mortality among 176 incinerator workers who were employed at least 1 year or more between 1920 and 1985 at a MSW incinerator in Sweden. The results revealed an excess mortality from cancer (oesophageal cancer RR = 2.84; stomach cancer RR = 1.27, rectal cancer RR = 2.52, lung cancer RR = 3.55, bladder cancer 1.98, malignant cerebral tumors RR = 2.77, hematopoietic cancer RR = 1.35) and nervous disease (RR = 1.33), circulatory disease (ischemic heart disease RR = 1.38), respiratory disease (asthma, bronchitis, emphysema RR = 1.62) and digestive disease (liver cirrhosis RR = 4.54). The excess was found to be high in workers with more than 40 years exposure.

Counter to the above study by Gustavsson [39], a retrospective study on 532 workers employed at two municipal waste incinerators in Rome did not reveal any excess of lung cancer [40]. Mortality from lung cancer was reduced in comparison with the general population and overall cancer mortality did not differ much from that of the general population. However, it was noted a 2.79 fold increased risk of mortality from gastric cancer among workers who had more than 10 years latency since first employment.

A similar study was conducted by Hours *et al.* [41]; they carried out a cross-sectional morbidity study for 102 workers employed at three French incinerators during 1996, matched for age with 94 male workers from other industrial activities. The exposed workers were categorised into three exposure groups based on their workplace: crane and equipment operators, furnace workers, and maintenance and effluent-treatment workers. The maintenance and effluent group encountered elevated relative risks for skin symptoms (RR = 4.85). An excess of daily cough was reported for the maintenance and effluent group (RR = 2.55) and for the furnace group (RR = 6.58).

Many epidemiologic studies dealing with waste management report limitations regarding a lack of good exposure data and the use of surrogate indirect measures which might lead to exposure misclassification [42–45] of the reasons for that is the unsuitableness of conducting an epidemiologic study based on experiments (not on observations) for ethical reasons.

4 CONCLUSION

Although there are a number of studies in the literature which provide information on emissions to air from waste treatment facilities, studies surveying the emissions to land or water are very limited in number. This does not mean that health effects due to exposure via water or soil are less significant; however, there are controls on food and water quality, which make any exposures through these pathways easier to avoid. Therefore, inhalation of emissions is the pathway, which is mostly assumed by epidemiological studies.

The greatest challenge emphasised in the current literature so far is the “confounding factors” which might not adequately be controlled in many studies such as ethnicity, gender, socio-economic or deprivation status, age, smoking/alcohol habits, medicinal drug use, occupational history, hazards from other sources, population mobility, long latency period of

some diseases, the pre-existing health of the people being studied, the wealth or poverty of the people, the availability of health or social care services and other present or historical sources of pollution.

It is known that adverse health impacts would be difficult to prove or supply with decent figures. The main conclusion of the review of the epidemiology literature is that the evidence of adverse health outcomes is controversial as they are insufficient/inadequate and hence inconclusive in providing fully convincing, rigorous epidemiological evidence for an association between waste treatment facilities and adverse health outcomes. It is a fact that future research into the health risks of waste management needs to overcome these current limitations.

It is, therefore, suggested that further collaborative epidemiological studies using a more rigorous approach along with an appropriate methodology which takes account of possible confounding factors are required. It is anticipated that this will benefit in improving a way of shaping public perspective through waste treatment facilities that underlies social values in waste management decision making.

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